

Grazing impacts on litter and roots: perennial versus annual grasses

E. MAPFUMO, M. A. NAETH, V. S. BARON, A.C. DICK AND D. S. CHANASYK

Authors are Post-doctoral Fellow and Professor in Department of Renewable Resources, University of Alberta, Edmonton, Alberta, Canada T6G 2H1, Research Scientist and research scientist with Lacombe Research Centre, Agriculture and Agri-Food Canada, 6000 C&E Trail, Lacombe, Alberta, Canada T4L 1W1, and Professor in Department of Renewable Resources, University of Alberta, Edmonton, Alberta, Canada T6G 2H1.

Abstract

Soil carbon (C) and nitrogen (N) storage in grasslands is a function of litter and root mass production. Research on how annual grasses compare with perennials for above ground and below ground mass production, and contributions to the soil C pool under pasture management is scarce. The objective of this research was to evaluate grazing intensity effects on litter and root mass, C and N pools of perennial grasses, smooth brome (*Bromus inermis* L.) and meadow brome (*Bromus riparius* Rhem.), and the annual grass, winter triticale (*X Triticosecale* Wittmack). Litter mass and C pool for the perennial grasses were greater than those for triticale. Litter C and N pools generally decreased with increased grazing intensity. Root mass was greater for the perennial grasses than for triticale at all grazing intensities. Meadow brome generally produced more root mass than smooth brome. Root C and N pools for triticale were 31 and 27%, respectively, of that for the perennial grasses. Estimated total C contribution (roots and litter) to the resistant soil organic C pool was 1.5 times greater for light compared to heavy grazing. Total C (litter + root) contribution for perennial grasses was 2.7 times greater than that for triticale. Perennial grasses provided a larger litter base and root system that promote greater storage of C in the soil compared with triticale.

Key Words: annuals, organic C, perennials, total N, sequestration

Grazing has a major impact on litter, roots, and soil characteristics. Litter herein refers to all dead (standing and fallen) plant material above the soil surface (Naeth 1988). Litter reduces soil erosion by reducing runoff and improves soil structure and fertility through addition of organic matter (Naeth 1988). Bare soils are more susceptible to raindrop impact and aggregate break down which can lead to surface sealing and increased erosion. Litter is especially critical at snowmelt and during intense rainfall events that can potentially remove large amounts of surface soil and nutrients (Chanasyk and Woytowich 1987). The amount of litter in a pasture is a function of forage growth, senescence, harvest and decomposition (Coleman 1992). Grazing affects plant characteristics primarily via biomass and litter removal. Standing and fallen litter mass and vegetation ground cover generally decrease while amount of bare ground increases with increased grazing

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Resumen

El almacenamiento de nitrógeno (N) y carbón © en los suelos de pastizal esta en función de la producción de mantillo y biomasa radical. Hay una escasez de investigación respecto a como los zacates anuales, comparados con los perennes, producen biomasa aérea y subterránea y las contribuciones a la reserva de C del suelo bajo praderas manejadas. El objetivo de esta investigación fue evaluar el efecto de la intensidad de apacentamiento en el mantillo y la biomasa de raíces de los zacates perennes, "Smooth brome" (*Bromus inermis* L.) y "Meadow brome" (*Bromus riparius* Rhem.) y de la especie anual de triticales invernal (*X Triticosecale* Wittmack) y en las reservas de C y N. La biomasa de mantillo y la reserva de C de los zacates perennes fue mayor que la del triticale. Las reservas de C y N del mantillo generalmente disminuyeron al aumentar la intensidad de apacentamiento. La biomasa de raíces fue mayor para los zacates perennes que para el triticale, esto se registró en todas las intensidades de apacentamiento. El "Meadow brome" generalmente produjo más biomasa de raíces que el "Smooth brome". Las reservas de C y N de las raíces del triticale fueron 31 y 27%, respectivamente, de las reservas registradas en los zacates perennes. La contribución total estimada de C (raíces y mantillo) a la reserva de C del suelo orgánico resistente fue 1.5 veces mayor para el apacentamiento ligero que para el apacentamiento fuerte. La contribución total de C (mantillo + raíces) de los zacates perennes fue 2.7 veces mayor que la del triticale. Los zacates perennes proporcionaron una mayor base de mantillo y sistema radical lo que promueve un mayor almacenamiento de C en el suelo comaprado con el triticale.

intensity (Naeth et al. 1991).

Grasslands have the ability to store substantial pools of soil C and N. Grasslands contain about 10% of the world C pool (Parton et al. 1995). In temperate regions grasslands may release as much as 40% of their C through respiration when they are cultivated and converted to grains and oilseeds (Burke et al. 1995). This narrows the soil carbon-to-nitrogen ratio (C:N) favoring the release of soil N through N mineralization processes (Wedin 1996). Perennial pastures and hay crops on the Canadian prairies are usually found as a 2 to 9 year sequence in a forage-cereal rotation (Entz et al. 1995). During this time they contribute to soil C and N pools and after being broken enhance yields of cereal crops (Campbell et al. 1990, Entz et al. 1995). The residue that remains goes into the litter pool and subsequently a portion is sequestered in soil. Initial break down results in large losses of organic C stored under grasslands and annual cultivation exacerbates the loss (Campbell et al. 1990).

The objective of this study was to quantify grazing management effects on litter and root mass, litter and root C and N pools in 2 perennial pastures, smooth brome grass (*Bromus inermis* L.) and meadow brome grass (*Bromus riparius* L.), and triticale (*X Triticosecale* Wittmack) when grazed at light to heavy grazing intensities.

Materials and Methods

Site Description and Meteorological Conditions

The study was conducted at Lacombe, Alberta (52°28'N; 113°45'W; 847 m), on a Typic Haplustoll derived from glaciolacustrine parent material. The upper 15 cm of soil averaged 15% clay, 34% silt, and 51% sand. Soil pH using distilled water was 5.4 and the sodium adsorption ratio was 0.2.

Total precipitation between April and October for 1993, 1994, 1995, and 1996 was 416, 530, 408, and 383 mm, respectively. The long-term normal annual precipitation (89-year average) was 362 mm. Mean air temperature between April and October ranged from 6.5 to 13.6° C in 1993, 7.7 to 17.1° C in 1994, 5.8 to 15.9° C in 1995 and 2.8 to 16.1° C in 1996. During the winter months (November to March) the total precipitation was 89 mm in 1993/1994, 30 mm in 1994/1995, 88 mm in 1995/1996, and 235 mm in 1996/97.

Experimental Design, Species and Grazing Treatments

Before being broken the site had been under extensive grazing management of 15-year old perennial grass pasture that was composed of smooth brome grass (*Bromus inermis* L.), quackgrass (*Elytrigia repens* L.) and Kentucky bluegrass (*Poa pratensis* L.). Cultivation of the plot area commenced in summer 1992 so that new species could be established in 1993. Data collection was initiated in 1994.

The experimental field plots were a randomized complete block design with 3 grazing treatments, 3 forage species and 4 replications as blocks. Each plot was 33 × 9 m and was subjected to 1 of 3 grazing treatments (heavy, medium, light). The upper two experimental blocks were east-facing on a 4 to 6% slope, while the other 2 blocks were on flat land.

Three forage species, 2 perennials and an annual, with potentially differing abili-

ties to produce litter were used. Carlton smooth brome grass (*Bromus inermis* L.) and Paddock meadow brome grass (*Bromus riparius* Rhem.) were the perennial species; Pika winter triticale (*X Triticosecale* Wittmack) was the annual species. Smooth brome grass is a rhizomatous species whereas meadow brome grass is a bunchgrass such that differences in root and litter masses could be expected. Prior to seeding the experimental area received a broadcast application of 8, 31, 31, and 5 kg ha⁻¹ of N, P₂O₅, K₂O, and S, respectively. This was followed by a light cultivation and packing. Smooth brome grass was seeded at 11.2 kg ha⁻¹ and meadow brome grass at 16.8 kg ha⁻¹. Spredor II alfalfa (*Medicago sativa* L.) was seeded with each grass at 1 kg ha⁻¹. Alfalfa had almost totally disappeared by the time measurements were taken in 1996. Perennial species were broadcast seeded with a Model HHBS-125 Handi-Spreading Lawn and Garden Seeder-Spreader. Seeding was followed by one pass with a diamond tooth harrow and one pass with a crowfoot packer. Each spring 100, 50, and 50 kg ha⁻¹ of N, P₂O₅ and K₂O, respectively were broadcast over the experimental area. Such amounts of fertilizer are not uncommon in intensively grazed pastures. Annual plots were rototilled to a depth of 10 cm and seeded to triticale at 135 kg ha⁻¹ in 2 passes with row spacing of 12.5 cm. A herbicide, 2-methyl-4-chlorophenoxyacetic acid (MCPA), was applied in spring each year at a rate of 600 g ai ha⁻¹ in 1994 to 1996 and at a rate of 900 g ai ha⁻¹ in 1997, to the triticale to control weeds. No herbicides were applied in perennial grass plots. Residue (all above ground plant material remaining after the last grazing of the previous season) was left in place until spring (end of April) seeding.

All plots were grazed with one-year old crossbred beef heifers. In 1993, unquantified light grazing was used to reduce forage biomass by approximately 50% on all plots. Beginning in 1994, up to 8 animals were placed on a treatment at one time, depending on the intensity of grazing desired. Water was constantly available to cattle so as not to disrupt grazing habits. Grazing events lasted usually less than 24 hours.

Forage height was used to define grazing intensity. Grazing started when forages reached a target maximum and ceased when a target minimum height was reached. Target heights varied among species and were set according to the species morphology, the desired amount of litter and the amount of bare ground deemed appropriate for that treatment.

Forage heights were determined using a diskmeter (Bransby et al. 1977) and the average of 10 disk heights was calculated. The diskmeter was used in this case to maintain consistency in assessing forage availability. For perennials, heavy, medium and light grazing was initiated at 13, 17, and 26 cm and stopped at 6, 5, and 7 cm, respectively when averaged over all grazing events within a year. Comparable figures for triticale were 11, 12, and 21 cm on entry and 3, 4, and 6 cm on exit of animals from heavy, medium and light intensity grazing. Perennials were grazed 7, 5, and 3 times and the annuals 4, 4, and 2 times for heavy, medium, and light grazings, respectively within the 3 study years. Averaged over 3 years the animal unit months per hectare (AUM ha⁻¹, based on approximately 450 kg animal) for perennials was 45.2, 24.4, and 19.6 AUM ha⁻¹, for heavy, medium and light grazing, respectively. For triticale this was 23.3, 13.5, and 9.8 AUM ha⁻¹, for heavy, medium, and light grazing, respectively.

Vegetation Sampling

Litter samples were collected in fall 1994, 1995, and 1996 and in spring 1995, 1996, and 1997. Spring sampling occurred in late April, prior to cultivation of annual plots when perennials had already begun growing. Fall samples were collected between end of September and early October, after the final grazing event each year. Samples were collected from 3 randomly selected sites within each plot in each replicate. All vegetation within a 0.05-m² quadrat was clipped at ground level and the surface soil was raked with a hand-fork to remove litter above the soil-mineral surface. All material collected was separated into live (any green material) and dead components before oven-drying at 60° C to determine litter mass (all dead material standing and fallen).

Carbon and Nitrogen Pools in Litter

Three litter samples taken from each plot were combined before determination of the carbon (C) and nitrogen (N) fractions in litter using the Leco Carbon Determinator (Model CN 2000, Leco Corp., St. Joseph, Minn.) and the Kjeldahl digestion method (McGill and Figueiredo 1993), respectively. The litter C and N pools were the product of litter mass (kg ha⁻¹) multiplied by percent C or N content. Litter C and N pools were compared among grazing levels, plant species, and years.

Root Sampling

Soil cores were taken in October 1996 and 1997 to a depth of 60 cm from 3 random locations in non-crown areas in each plot in each replicate using a hydraulically powered sampler with a 5.1-cm diameter probe. Soil cores were separated into depth segments of 0 to 5, 5 to 15, 15 to 30, and 30 to 60 cm and bulked by depth for each plot. Samples were transported from the field immediately, spread in shallow pans and dried at room temperature in a forced-air dryer. Roots were separated from soil with a hydropneumatic elutriation system (Smucker et al. 1982). No attempt was made to differentiate between live and dead roots. Root material was collected on a fine-mesh screen and transferred by washing on to Whatman No. 4 filter papers 9 cm in diameter. The papers and roots were dried in a forced air oven at 85° C for at least 24 hours and then weighed. Root mass per hectare for each sampling depth was calculated from the sample core size. Due to small sample sizes, the samples for each depth within each replicate were combined and ground through a cyclone mill (UDY Corporation, Boulder, Colo.) to pass a 0.5-mm screen. Total C and total N were measured with a Leco Carbon Determinator (Model CN 2000, Leco Corp., St. Joseph, Minn.) and the Kjeldahl digestion method (McGill and Figueiredo 1993), respectively. Root masses for the 0 to 15 cm and 0 to 60 cm were examined to determine if there were differences among grazing levels and species in the total root mass (0 to 60 cm) and in the amount of roots found in the plough layer (0 to 15 cm).

Statistical Analyses

Statistical analyses was conducted using the SAS statistical package generalized linear models procedure (SAS Institute 1989). Analyses of litter mass, litter C pool, litter N pool, root mass, root C pool, and root N pool were conducted using a split-split-plot design repeated across years. Grazing intensity and species were main plot effects tested for significance using replicate within (grazing X species) as the error term, years were a subplot effect tested with replicate within years as the error term, and the interactions were sub-subplot effects tested with the residual error (Steel and Torrie 1980). Where the F-test indicated a significant ($P \leq 0.05$) effect, means were separated by calculation of least significance difference (LSD) using the appropriate error mean squares (Gomez and Gomez 1984).

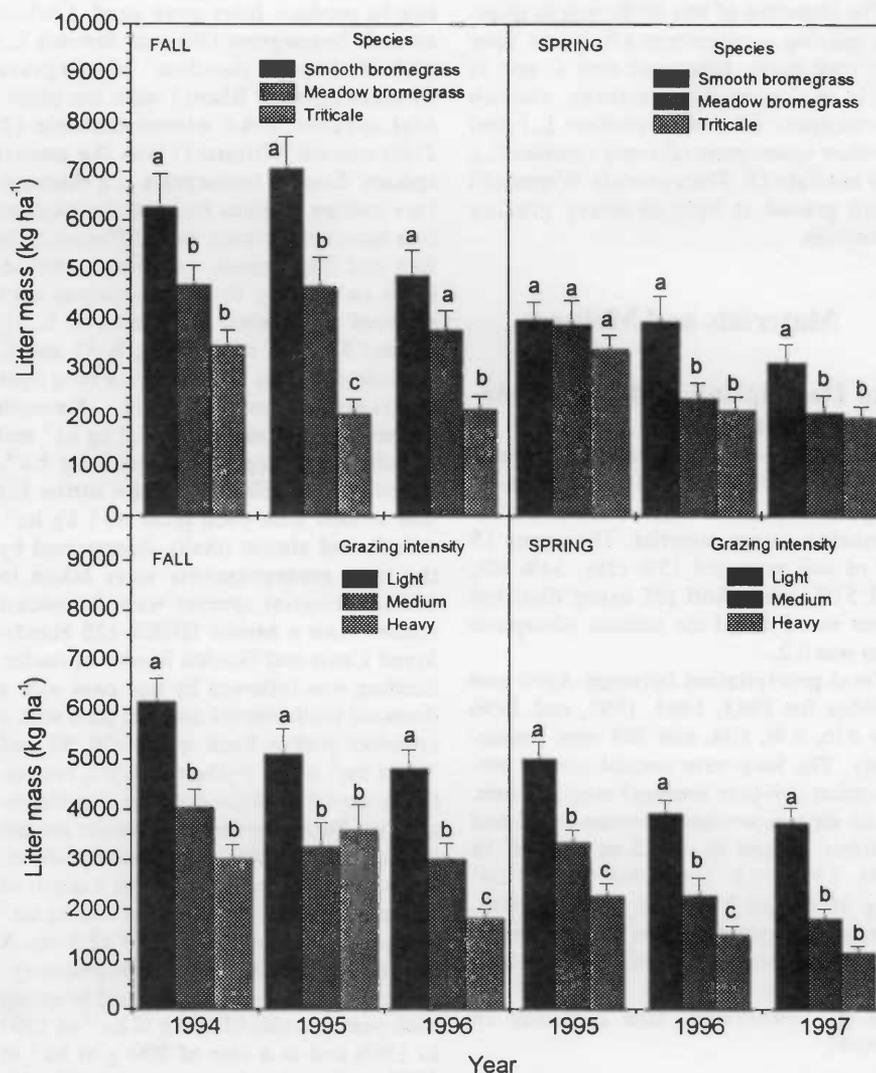


Fig. 1. Litter mass (kg ha^{-1}) in fall and spring under different species, years, and grazing intensities.

Results

Litter Mass, Carbon, and Nitrogen Pools

Litter mass in fall was greater for perennial grasses than for triticale whereas in spring litter mass for smooth bromegrass was greater than that for other species (Fig. 1). Averaged across years, litter mass in fall for triticale was only 48% that for perennials whereas in spring it was 78%. In both spring and fall, litter mass decreased with increasing grazing intensity (Fig. 1). Heavy and medium grazing had average fall litter masses 53 and 64% that for light grazing, respectively. However, in spring, litter masses for heavy and medium grazing were 39 and 59% that for light grazing, respectively.

Differences in litter C pool among species were evident in fall of all years. In

all years the C pool in fall was greatest for smooth bromegrass and smallest for triticale whereas in spring litter C pool differences among species were low and inconsistent (Fig. 2). In spring, litter C for triticale across grazing intensities was 72% of that for perennials, and litter masses for heavy and medium grazing were, respectively, 36 and 57% of that for light grazing. In general, the litter C pool in both fall and spring decreased with increased grazing intensity (Fig. 2). In 4 out of 6 cases the C pool of the heavy grazing treatment was less than 50% of that for the light grazing treatment.

The litter N pool in 2 out of 3 years was greater for perennials than for triticale, with smooth bromegrass having the largest N pool of all forages and triticale having the smallest (Fig. 3). Litter N pool in spring for triticale was 71% of that for

perennials. In fall, the heavy grazed treatment had a smaller N pool in litter than the light grazed treatment whereas the N pool of the medium grazed treatment was intermediate. Furthermore, heavy and medium grazing litter masses were 41 and 62% of that for light grazing, respectively.

Litter C:N ratios ranged between 13:1 and 16:1 and were similar among forages in both fall and spring (Fig. 4). The litter C:N ratio in fall generally decreased with increased grazing intensity. For spring, litter C:N ratios for heavily grazed forages were lower than that for lightly grazed forages in 2 out of 3 years (Fig. 4).

Root Mass, Carbon and Nitrogen Pools

Trends among grazing intensities for accumulated root mass to 15 cm or 60 cm (Table 1) were similar. The majority of root material (52 to 60%) existed in the upper 15 cm. Root mass of triticale was always substantially less than that of the perennial grasses throughout the soil profile (Table 1). Root mass of the perennial grasses was 2.6 to 5.2 times that of triticale within grazing levels.

Grazing intensity affected root mass of the species differently (Table 1). Over the 60-cm depth and within the 0 to 15-cm depth root mass of triticale was unaffected by grazing intensity. Root mass of smooth brome grass maximized at the medium grazing intensity for the 0 to 60-cm depth interval and decreased between medium and light grazing intensities for the 0 to 15-cm depth interval (Table 1). By contrast root mass of meadow brome grass increased from medium to light grazing intensity for the 0 to 60-cm depth interval and for the 0 to 15-cm depth interval (Table 1).

Perennial grasses always had signifi-

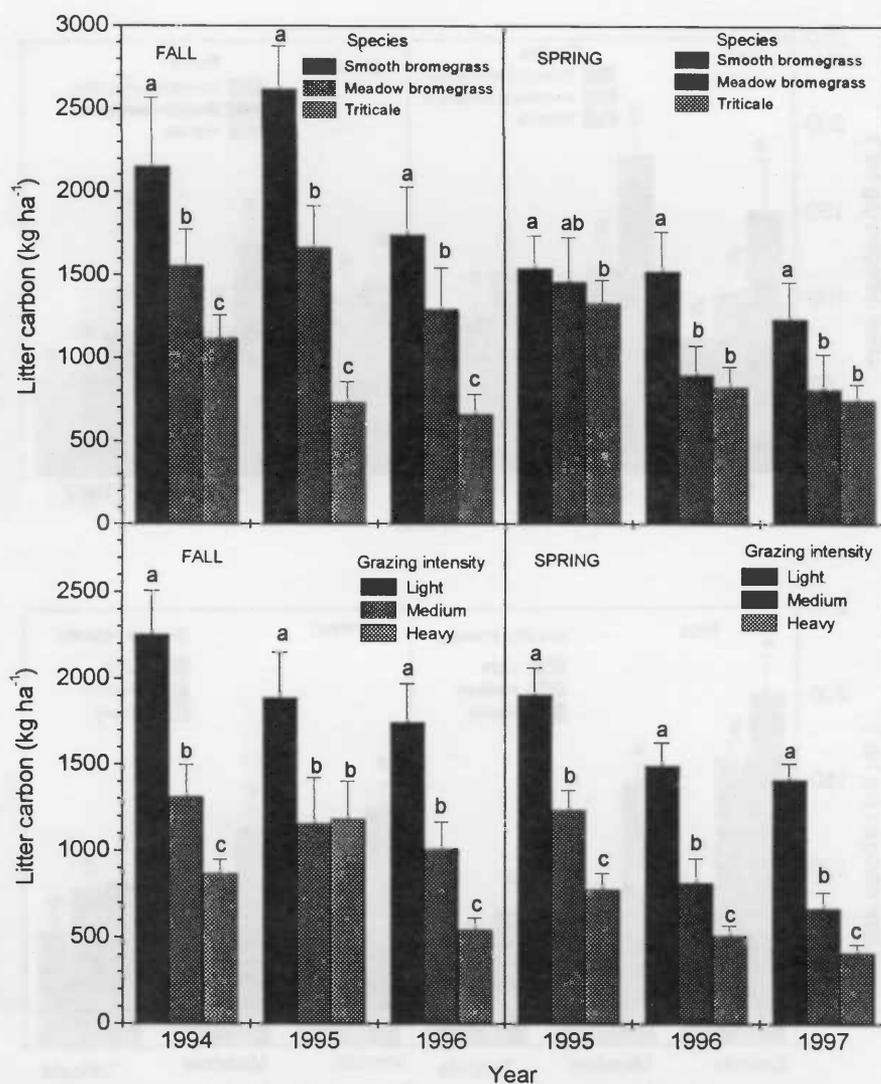


Fig. 2. Litter carbon (kg ha^{-1}) in fall and spring under different species, years and grazing intensities.

cantly higher root C and N pools than triticale (Table 2). Averaged over grazing level triticale had 25 to 30% of the root C

and 28 to 32% of the root N pool size of perennial grasses. For smooth brome grass the C and N pools to 60 cm depth were greatest in the medium grazing intensity. The root C pool for meadow brome grass under heavy grazing intensity was 67% of that under light grazing intensity. The root N pool for meadow brome grass increased up to the light grazing intensity with the heavy grazing intensity 75% of the light. The widest range among species for C and N pool size occurred at the light grazing intensity. At this grazing intensity, triticale and smooth brome grass were 20 and 81%, respectively, of meadow brome grass for the root C pool, and 28 and 85%, respectively, of meadow brome grass for the root N pool.

Significant variation was observed among species and from year to year for root C:N ratio (data not shown). There

Table 1. Root mass (kg ha^{-1}) in the surface (0 to 15 cm) segment and for the whole profile (0 to 60 cm) under 3 grass species subjected to three grazing intensities averaged over 2 years.

Species	Grazing intensity		
	Light	Medium	Heavy
----- Root mass (kg ha^{-1}) -----			
Surface (0-15 cm)			
Smooth brome grass	1390 ± 164 bB	2020 ± 280 aA	1720 ± 275 aA
Meadow brome grass	2610 ± 402 aA	2090 ± 222 aB	1540 ± 382 aC
Triticale	530 ± 68 cA	800 ± 209 bA	480 ± 125 bA
Total (0-60 cm)			
Smooth brome grass	2840 ± 263 bB	3520 ± 243 aA	3120 ± 295 aB
Meadow brome grass	4890 ± 539 aA	3730 ± 331 aB	3320 ± 728 aB
Triticale	940 ± 82 cA	1320 ± 246 bA	810 ± 133 bA

Values presented are means ± standard errors for 8 replicates.

Means within column and depth increment followed by different lower case letters indicate significant difference among species within grazing intensities (Least Squares Means, $P \leq 0.05$).

Means within row and depth increment followed by upper case letters indicate significant difference among grazing intensities within species (Least Squares Means, $P \leq 0.05$).

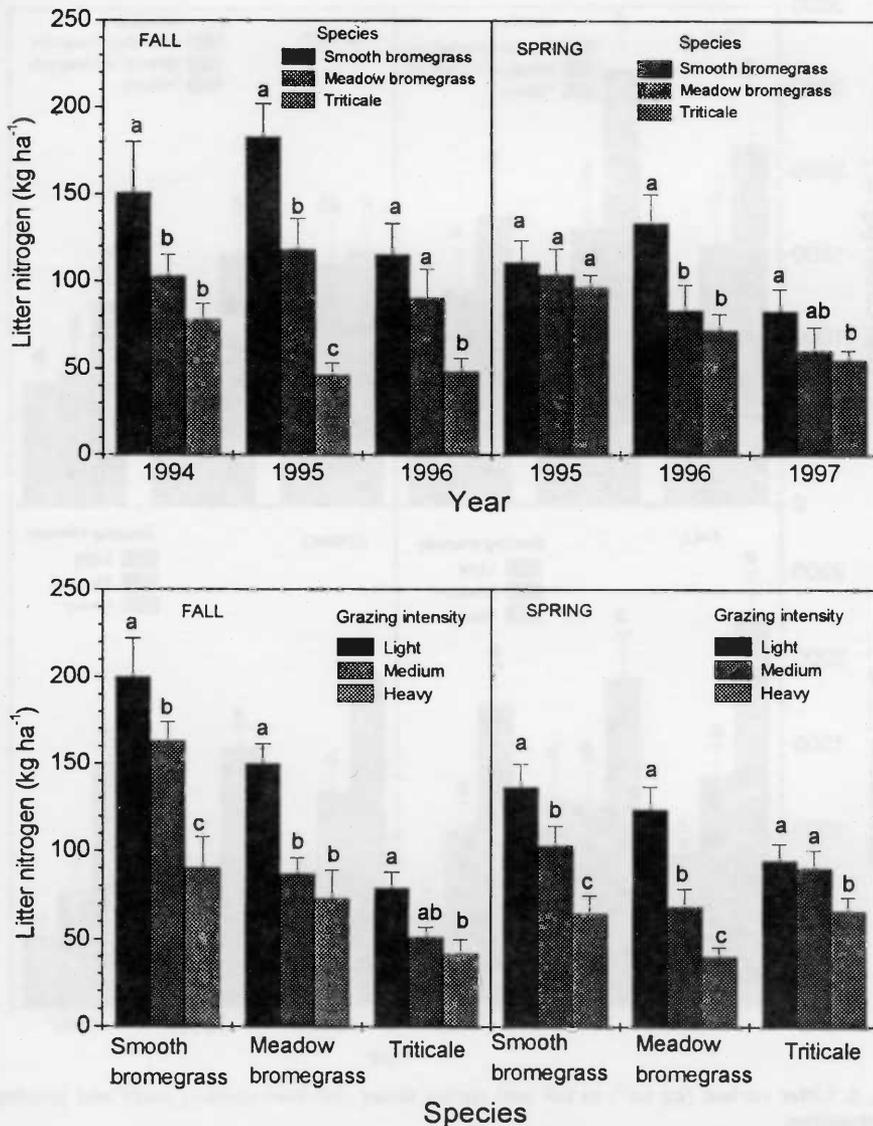


Fig. 3. Litter nitrogen (kg ha^{-1}) in fall and spring under different species, years and grazing intensities.

was a trend for triticale to have a lower C:N ratio than smooth and meadow bromegrass, but the difference was only significant in 1997. The year-to-year dif-

Table 2. Carbon and nitrogen pools in roots of annual and perennial grasses subjected to light, medium and heavy grazing.

Species	Grazing intensity		
	Light	Medium	Heavy
	----- Carbon pool (kg ha^{-1}) -----		
Smooth bromegrass	1190 ± 117 bB	1490 ± 117 aA	1270 ± 123 aB
Meadow bromegrass	1960 ± 214 aA	1450 ± 107 aB	1320 ± 271 aB
Triticale	400 ± 34 cA	460 ± 86 bA	310 ± 51 bA
	----- Nitrogen pool (kg ha^{-1}) -----		
Smooth bromegrass	42.2 ± 3.7 bB	59.7 ± 8.3 aA	45.9 ± 3.9 aB
Meadow bromegrass	66.2 ± 7.7 aA	54.1 ± 8.6 aAB	49.8 ± 15.1 aB
Triticale	15.4 ± 1.6 cA	19.1 ± 3.6 bA	13.0 ± 1.7 bA

Values presented are means ± standard errors for 8 replicates.

For each pool means within each column followed by different lower case letters indicate significant difference among species within grazing intensity (Least Squares Means, $P \leq 0.05$).

For each pool means within each row followed by upper case letters indicate significant difference among grazing intensity within species (Least Squares Means, $P \leq 0.05$).

ferences in C:N ratio reflected variations in the root-N concentration. Overall the root C:N ratio ranged between 22:1 and 36:1.

Discussion

Litter mass, Litter C and N Pools

The benefits of litter through improved soil structure and infiltration as well as decreased raindrop impact, runoff and evaporation have been reported widely (Willms et al. 1986, Naeth et al. 1990). In our study, accumulations of litter were strongly influenced by grazing intensity as well as plant species and year. The reduced amount of litter on heavily grazed treatments compared to lightly grazed ones may be attributed to greater removal of green herbage through grazing. It is also possible that heavy grazing may have accelerated litter decay through trampling compared with medium and light grazing. This result has been reported for other grasslands (Rhoades et al. 1964) including grazed lands in Alberta (Naeth et al. 1990, Dormaar and Willms 1992).

Treatment differences in the C and N pool of litter were most clearly expressed in fall. Responses in litter C to grazing intensity were clear with light grazing always providing more C than heavy. The seasonal differences in the carbon content of litter under each grazing intensity may be partly due to freezing effects (loss of cell contents) over winter. Furthermore, a concurrent study conducted on the same plots indicated that annual runoff was dominated by snowmelt-induced runoff, averaging 98, 84 and 86% for the light, medium and heavy grazing, respectively (Gill et al. 1998). This implies that the physical loss of litter could have occurred during snowmelt in spring resulting in reduced litter and litter C pools compared with pools in fall of the previous year.

Quality of litter is reflected in the C:N ratio and lignin contents. Ultimately soil C and C:N ratio will reflect litter and root characteristics that influence decomposition rates (Wedin and Tilman 1990). Because the rate of microbial decomposition is related to the C:N ratio, it is expected that the proportion of C and N reaching the soil would be proportionately less than that indicated by C and N pools from litter. The lignin content controls the split of litter into structural and metabolic material. Most of the structural material (70%) with high lignin is stabilized in the soil while very little is found in microbial biomass (Parton et al. 1987).

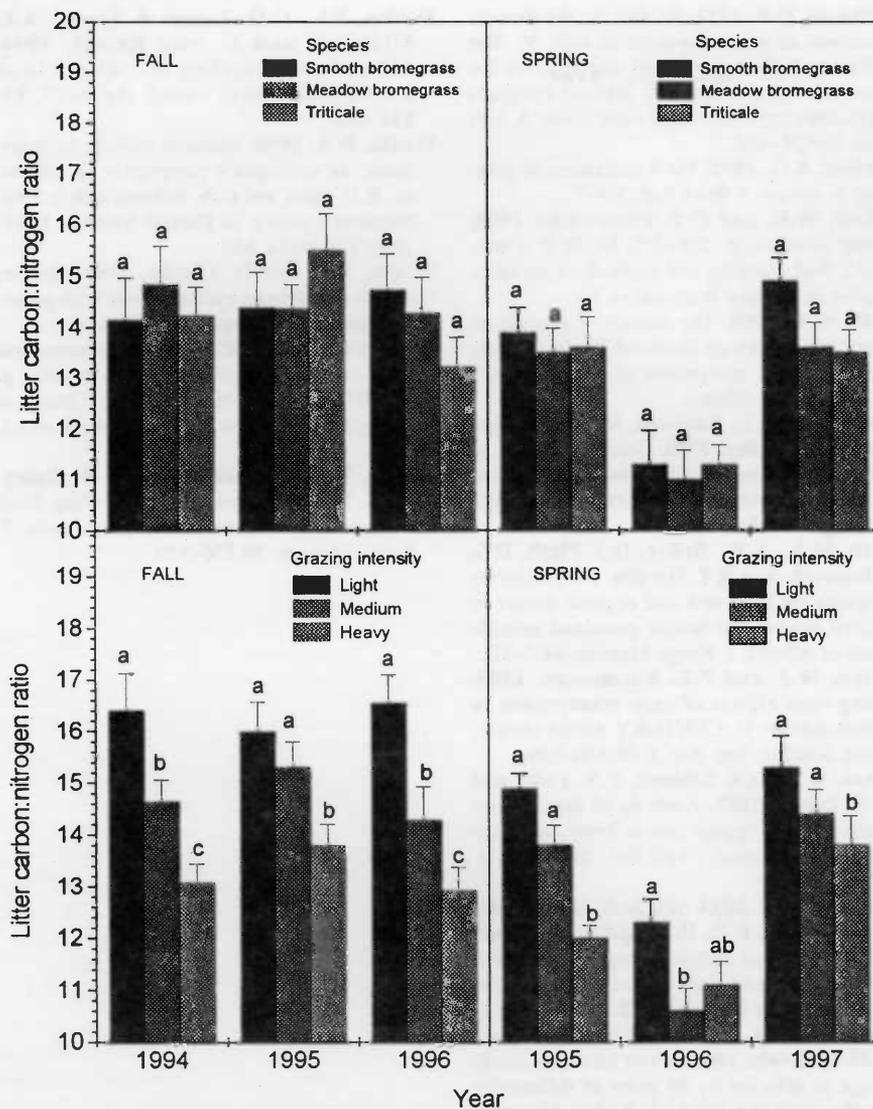


Fig. 4. Litter carbon:nitrogen ratio in fall and spring under different species, years and grazing intensities.

Root Mass, Root C and N Pools

Root mass is expected to decrease with increased grazing intensity (Briske 1991). Meadow bromegrass appeared to follow this trend for root mass with increased grazing intensity. However, smooth bromegrass under light grazing produced a lower root mass than that under moderate grazing. The difference in responses between the bromegrass species for grazing intensity is difficult to explain. Matches (1992) cited studies of other species where grazing intensity did not affect root mass (as for triticale) and Harker and O'Sullivan (1993) observed that quackgrass was stimulated to produce root material relative to untreated controls when thinned by herbicide application. Their conclusion was that rhizomatous species (like quackgrass and smooth bromegrass) might suffer from intraspecific competition for root development, when

allowed to grow uncontrolled.

In this study the perennials consistently had 3.7 times the C and 3.3 times the N pools in the root material compared with triticale, possibly because perennial roots include both live and dead roots whereas triticale had mostly live roots. Turnover of root material is an important consideration in determining the annual contribution of roots to the soil C and N pool. Roots of triticale were produced annually, and annual cultivation would increase the rate of decomposition of dead roots in the surface layer. The turnover of perennial grass roots is more difficult to assess because management level can influence lifespan of roots. Grazing, cutting, and fertilizer application, tends to shorten the average turnover period (Whitehead 1995). Lifespan of roots may vary from 4 to 6 weeks to up to several years in shortgrass prairie (Whitehead 1995). Decomposition

rates are greater in root material with low C:N ratios (Whitehead 1995). In a similar environment Walley et al. (1996) reported an average root turnover rate of 1 year for alfalfa and meadow bromegrass.

Litter and Roots Contribution to the Stable Soil C Pool

According to Van Veen and Paul (1981), 50% of the litter produced annually on native grassland enters the soil. In cereal grasses 80% of the litter that enters the soil is easily decomposed while the remaining 20% is more resistant to decomposition (Jenkinson 1977; Parton and Rasmussen 1994). An estimated 63% of the root C is easily degradable with the remaining 37% forming part of the resistant soil organic matter (Van Veen and Paul 1981). Using these percentages the estimated C contributions from litter and roots resulted in total C contributions to resistant organic C of 1,210, 1,550, and 1,810 kg C ha⁻¹ for heavy, medium, and light grazing, respectively. Over the same period (i.e. 3 years), the total C contribution to resistant soil organic C was 2,608 kg C ha⁻¹ for perennials compared to only 962 kg C ha⁻¹ for triticale. However, these estimates are simplified because they did not take into account factors that make C contribution estimation more complicated, such as yearly cultivation of annual species and lignin concentration.

Within the top 30 cm, total soil organic C did not change significantly over the period of study (1994 to 1996) and was not affected by grazing treatments (Baron et al. 1999). Average soil total C concentrations over the 0 to 15-cm depth interval were 5.3% in 1994 and 5.4% in 1997, which is equivalent to 74 and 75 Mg ha⁻¹ C, respectively. This observation is consistent with findings on Black Chernozems with high organic matter content (Campbell et al. 1991).

In conclusion, increased grazing intensity resulted in smaller litter C and N pools. A similar trend was observed for root C and N pools. Both litter and root mass, and litter C pool for perennials were greater than that for triticale. Thus growing perennials would provide a greater litter base, greater litter C and root C pools than annuals. Amidst concern about global warming, growing perennial species may potentially enhance C sequestration and reduce net emission of carbon dioxide from agricultural ecosystems.

References

- Baron, V.S., E. Mapfumo, M.A. Naeth, and D.S. Chanasyk. 1999.** Sustainable grazing systems for perennial and annual forages on sloped lands. Canada-Alberta Environ. Sustainable Agr. Agreement Final Rep. 161 pp.
- Bransby, D.I., A.G. Matches, and G.F. Krause. 1977.** Disk meter for rapid estimation of herbage yield in grazing trials. *Agron. J.* 69:393–396.
- Briske, D.D. 1991.** Developmental morphology and physiology of grasses, p. 85–108. *In:* R.K. Heitschmidt and J.W. Stuth (eds.), *Grazing management: an ecological perspective*. Timber Press, Portland, Ore.
- Burke, I.C., W.K. Laurenroth, and D.P. Coffin. 1995.** Soil organic matter recovery in semi arid grasslands: implications for the Conservation Reserve Program. *Ecol. Appl.* 5:793–801.
- Campbell, C.A., R.P. Zentner, H.H. Janzen, and K.E. Bowren. 1990.** Crop rotation studies on the Canadian prairies. Research Branch, Agriculture Canada. Pub. No. 1841/E.
- Campbell, C.A., K.E. Bowren, M. Schnitzer, R.P. Zentner, and L. Townley-Smith. 1991.** Effect of crop rotations and fertilization on soil organic matter and some biochemical properties of a thick Black Chernozem. *Can. J. Soil Sci.* 71:377–387.
- Chanasyk, D.S. and C.P. Woytowich. 1987.** Sediment yield as a result of snowmelt runoff in the Peace River Region. *Can. Agr. Eng.* 29:1–6.
- Coleman, S.W. 1992.** Plant-animal interface. *J. Prod. Agr.* 5:7–13.
- Dormaer, J.F. and W.D. Willms. 1992.** Water-extractable organic matter from plant litter and soil of rough fescue grassland. *J. Range Manage.* 45:152–158.
- Entz, M.H., W.J. Bullied, and F. Katema-Mupondwa. 1995.** Rotational benefits of forage crops in Canadian Prairie cropping systems. *J. Prod. Agr.* 8:521–529.
- Gill, S.L., M.A. Naeth, D.S. Chanasyk, and V.S. Baron. 1998.** Runoff and sediment yield from snowmelt and rainfall as influenced by forage type and grazing intensity. *Can. J. Soil Sci.* 78: 699–706.
- Gomez, K.A. and A.A. Gomez. 1984.** Statistical procedures for agricultural research, 2nd ed. John Wiley and Sons, New York.
- Harker, K.N. and P.A. O'Sullivan. 1993.** Herbicide comparisons on quackgrass (*Elytrigia repens*) within different crop competition tillage conditions. *Weed Sci.* 41:94–99.
- Jenkinson, D.S. 1977.** Studies on the decomposition of plant material in soil. V. The effects of plant cover and soil type on the loss of carbon from ¹⁴C labeled ryegrass decomposing under field conditions. *J. Soil Sci.* 28:424–434.
- Matches, A.G. 1992.** Plant responses to grazing: A review. *J. Prod. Agr.* 5:1–7.
- McGill, W.B. and C.T. Figueiredo. 1993.** Total nitrogen, p. 201–212. *In:* M.R. Carter (ed.), *Soil sampling and methods of analysis*. Lewis Publishers, Boca Raton, Fla.
- Naeth, M.A. 1988.** The impact of grazing on litter and hydrology in mixed prairie and fescue grassland ecosystems of Alberta. Ph.D. Diss. Univ. of Alberta.
- Naeth, M.A., R.L. Rothwell, D.S. Chanasyk, and A.W. Bailey. 1990.** Grazing impacts on infiltration in mixed prairie and fescue grassland ecosystems of Alberta. *J. Range Manage.* 70:593–605.
- Naeth, M.A., A.W. Bailey, D.J. Pluth, D.S. Chanasyk, and R.T. Hardin. 1991.** Grazing impacts on litter and soil organic matter in mixed prairie and fescue grassland ecosystems of Alberta. *J. Range Manage.* 44:7–12.
- Parton, W.J. and P.E. Rasmussen. 1994.** Long-term effects of crop management in wheat-fallow: II. CENTURY model simulations. *Soil Sci. Soc. Am. J.* 58:530–536.
- Parton, W.J., D.S. Schimel, C.V. Cole, and D.S. Ojima. 1987.** Analysis of factors controlling soil organic matter levels in Great Plains Grasslands. *Soil Sci. Soc. Am. J.* 51:1173–1179.
- Parton, W.J., J. M.O. Scurlock, D.S. Ojima, D.S. Schimel, D.O. Hall, and S. Copegam. 1995.** Impact of climate change on grassland production and soil carbon worldwide. *Global Change Biol.* 1:13–22.
- Rhoades, E.D., L.F. Locke, H.M. Taylor, and E.H. McIlvain. 1964.** Water intake on sandy range as affected by 20 years of differential cattle stocking rates. *J. Range Manage.* 17:185–190.
- SAS Institute, 1989.** SAS/STAT user's guide Version 6, 4th Ed. SAS Institute, Cary, N.C.
- Smucker, A.J.M., S.L. McBurney, and S.K. Srivastava. 1982.** Quantitative separation of roots from compacted soil profiles by hydropneumatic elutriation system. *Agron. J.* 74:500–503.
- Steel, R.G.D. and J.H. Torrie. 1980.** Principles and procedures of statistics: a biometrical approach. 2nd edition. McGraw-Hill, N. Y.
- Van Veen, J.A. and E.A. Paul. 1981.** Organic carbon dynamics in grassland soils. I. Background information and computer simulation. *Can. J. Soil Sci.* 61:185–201.
- Walley, F.L., G.O. Tomm, A. Matus, A.E. Slinkard, and C. van Kessel. 1996.** Allocation and cycling of nitrogen in an alfalfa-bromegrass sward. *Agron. J.* 88: 834–843.
- Wedin, D.A. 1996.** Nutrient cycling in grasslands: an ecologist's perspective, p. 29–44. *In:* R.E. Joost and C.A. Roberts (eds.), *Proc. Nutrient Cycling in Forage Systems Conf.* 1996, Columbia, Mo.
- Wedin, D.A. and D. Tilman. 1990.** Species effects on nitrogen cycling: a test with perennial grasses. *Oecologia* 84:433–441.
- Whitehead, D.C. 1995.** Amounts, sources and fractionation of organic nitrogen in soils, p. 82–107. *In:* D.C. Whitehead (ed.), *Grassland nitrogen*. CAB International, Wallingford, U.K.
- Willms, W.D., S. Smoliak, and A.W. Bailey. 1986.** Herbage production following litter removal on Alberta native grasslands. *J. Range Manage.* 39:536–540.